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mechanism

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Abstract

The effect of different crystal phases, i.e. spinel phase (CuMn₂O₄) and amorphous phase (Cu_{0.2}MnO_x), was explored in Cu-Mn catalytic ozonation of toluene. The toluene removal efficiency followed the order of Cu_{0.2}MnO_x (91.2%) > CuMn₂O₄ (74.5%) > commercial catalyst Cu_{0.3}MnO_x (70.3%) in 130 min, and the higher CO₂ yield (67.6%) could be also observed using Cu_{0.2}MnO_x. In order to investigate the effect of phases on the toluene degradation pathway, the intermediates and byproducts were identified by DRIFTS, GC-MS, and TOF-SIMS. No obvious difference was observed in the distribution of byproducts, except for the quantities, suggesting the discrepancy of oxidation rate. On the other hand, the catalysts were characterized before and after the ozonation process by TEM, BET, XPS, XRD, EPR, TGA, and TPR. It was proposed that for amorphous catalysts, the oxygen vacancy (Vo) helped the chemisorption of toluene, and adjacent Mn reacted as the main active site for the ozonation process. While, the redox pair of Cu⁺/Mn⁴⁺ and Cu²⁺/(Mn³⁺, Mn²⁺) in the spinel phase plays an important role in the generation of oxygen vacancies for O₃ decomposition.

Keywords: Cu-Mn bimetallic oxide; Crystal phase; Catalytic ozonation; Reaction mechanism; Degradation pathway

1. Introduction

Volatile organic pollutants (VOCs) are the key precursors for urban photochemical smog, ozone (O₃), and atmospheric PM2.5, resulting in air pollution problems [1-3]. Therefore, VOCs abatement technologies have been widely studied. Among them, catalytic ozonation is recognized as an effective method for the low-temperature removal of VOCs [4, 5]. The removal efficiency has been highly investigated [6, 7]. The experimental parameters, such as humidity [6], temperature [8] and O₃ concentration [9], etc, have been regards as important factors. In addition, the selection of catalyst was also important for the ozonation process. MnO_x [10-14] is one of the popular catalysts used in catalytic ozonation of VOCs due to its outstanding performance in the abatement of VOCs and O₃. The performance of catalytic ozonation could be further improved when other metals are added into

MnO_x, such as Cu [15], Ce [16], etc. Jing Liu et al. [15] synthesized a series of copper doped manganese oxide octahedral molecular sieves (Cu-OMSx-T) with different Cu/Mn ratios and reported the best ratios of Cu/Mn was 0.5, and the improved performance was considered to be the synergistic effect between Cu and Mn (Cu²⁺ + Mn³⁺ \rightleftharpoons Cu⁺ + Mn⁴⁺) [14]. Considering the wide application of Cu-Mn oxides in catalytic ozonation of VOCs, it is worth further investigating the mechanism of the ozonation process, especially the effect of different Cu-Mn crystal phases.

Generally, the ozonation process over catalysts surface includes the adsorption of pollutant and O_3 , ozone activation to generate reactive oxygen species (O^- , O_2^- , $O_2^{2^-}$) for VOCs oxidation, and oxidation process [17]. The characteristic of catalysts plays an important role in the above-mentioned process, especially the different phases of catalyst [18, 19]. The different phases of Cu-Mn oxides can be controlled through various preparation methods, exhibiting different catalytic oxidation performances [20]. It is observed the well-distributed active sites of Cu-Mn and better redox properties using the amorphous sample resulting in excellent catalytic reactivity [21]. Chi-Woong Ahn et al. [22] found that the transformation of the catalyst crystal phase (from amorphous to crystalline phase) could strongly reduce the catalytic activity of the catalyst. Hong Chen et al. [23] synthesized amorphous Cu-Mn catalysts and reported good catalytic activity owing to the coexistence between crystallized CuO phase and amorphous Cu-Mn-O. Yu Wang et al. [24] reported that the interfacial structure of mixed phases (CuO and MnxOy) can induce the formation of Cu²⁺-O²⁻-Mn⁴⁺, as well as improved the reducibility. However, the effects of various crystal and morphologies on catalytic ozonation of VOC are still unclear, especially the degradation pathway and the synergetic effect of Cu and Mn in the different crystal phases.

Hence, in this paper, the samples with spinel and amorphous phases were prepared, respectively, and they are compared to a commercial amorphous Cu-Mn sample ($Cu_{0.3}MnO_x$) in catalytic ozonation of toluene. The performances were evaluated by toluene conversion, CO_2 yield, CO yield, byproduct composition, and O_3 abatement. The corresponding toluene degradation pathway was studied using DRIFTS, GC-MS. In addition, in order to analyze the role of the crystal phase during the reaction, the fresh and the used catalysts were characterized through SEM, BET, XRD, TPR, EPR, TOF-SIMS, TGA, and XPS.

2. Experiment

2.1 Catalyst preparation

The Cu_{0.2}MnO_x and CuMn₂O₄ catalysts were prepared by redox and co-precipitation methods, respectively. The detailed preparation process was described in our previous work [20]. The commercial catalyst was bought from Jining Yongbang Machinery Co. LTD and denoted as Cu_{0.3}MnO_x.

2.2. Catalyst characterization

The special surface area of the catalyst was determined according to the nitrogen adsorption and desorption isotherm, which is measured at -196°C on a Micromeretics ASAP 2460 instrument after 5 h of venting at 200°C. The pore size distribution was determined using the Barrett-Joyner-Halenda (BJH) method.

X-ray diffraction (XRD) analysis of all catalysts was performed on PNAlytical X'Pert PRO powder diffractometer using the Cu K α radiation at 40 kV and 40 mA. The 2 θ angle ranged from 10° to 80° with a step of 0.0167°.

The morphologies of catalysts were detected by scanning electron microscopy (SEM) images and high resolution transmission electron microscope (HRTEM) images recorded on a Hitachi S-4700 apparatus (II) and FEI Tecnai F30.

Inductively coupled plasma-optical emission spectroscopy (ICP-OES) was performed to obtain the metal elemental analysis of the samples (Cu and Mn) with the Agilent 720ES.

Electron paramagnetic resonance (EPR) experiments were performed with Bruker EMXPLUS (Germany).

Hydrogen temperature-programmed reduction (H2-TPR) was carried on the Chemisorb 2920 chemical adsorption analyzer (Micromeritics, USA). The catalysts were pretreated by pure He flows at 150°C (30 ml/min) for 1 h, and then heated with a rate of 5°C/min up to 500°C by using 5 vol.% H_2/Ar (50 ml/min) flow.

Thermogravimetric analysis and differential thermal analysis (TGA/DTA) were performed using the dried catalysts at a heating rate of 10°C/min up to 790°C with the use of Discovery TGA 5500.

X-ray photoelectron spectroscopy (XPS) was recorded with Thermo Scientific K-Alpha. The binding energies (BE) were calibrated using the peaks of C 1s at 284.8 eV. Peak fitting was processed with CasaXPS.

Gas chromatography-mass spectrometer (GC-MS) was used to identify the by-products on the Agilent 7890A-5975C instrument, which is equipped with a DB-5MS column.

Time of flight secondary ion mass spectrometry (TOF-SIMS) was carried out to analyze the surface composition of the sample with TOF-SIMS5 spectrometer (ION-TOF GmbH Germany) using 30 keV Bi³⁺ primary ions (pulsed current 0.1 pA).

The diffuse reflectance infrared Fourier transform spectroscopy (DRIFTS) was applied to detect the changes of functional groups on the surface of catalysts during the reaction. The instruments used in the experiment were Nicolet IS50 FTIR and Harrick *in situ* reaction chamber (HVC-DRM) of Thermo Fisher Company. The scanning time was set to 256 scans and the resolution was 8 cm⁻¹.

2.3. Catalytic activity evaluation

The toluene catalytic performance of the sample was evaluated in a fixed bed flow continuous reactor at atmospheric pressure. The schematic diagram of the experimental setup is shown in Fig. S1. About 200 mg of samples (40-60 mesh) was loaded in the quartz reactor (i.d. = 8 mm) and fixed in position with quartz wool on both ends of the catalyst. The simulated reaction gas contains ~160 ppm toluene introduced by micro-injector, O₃, and dry air. The flow rate and O3 concentration were adjusted by mass flow controllers (MFC) and fixed at 700 ml/min and 2100 ppm, respectively. The performance was evaluated at 100°C for 130 min. The concentration of toluene, CO, and CO₂ in outlet gas were detected by online gas chromatography (GC-9790H Fuli Gas Chromatography) equipped with thermal conductivity detector (TCD) and flame ionization detector (FID), and the columns of Restek Shin Carbon ST/Silco, HP NOC 80/100 micro-packed and capillary column Cp-Wax 52 CB25 m, ؼ 0.25 mm were installed. Toluene removal rate (TRR), CO yield (XCO) and CO2 yield (XCO₂), and others (XO) were calculated in supporting information.

3. Results and Discussions

3.1. Characterization of samples

 $Cu_{0.2}MnOx$ and $CuMn_2O_4$ catalysts were successfully synthesized (Fig. S2), and the corresponding catalyst characterization was in agreement with our previous report [20], exhibiting the phase of spinel $CuMn_2O_4$ and amorphous phase of $Cu_{0.2}MnO_x$. In order to further reveal the ozonation process of amorphous catalyst, a commercial catalyst was chosen with amorphous phase for comparison. As for $Cu_{0.3}MnO_x$, it was observed that the two wide diffraction peaks located in the 2 theta regions of 34-39° and 62-71°, indicating that they were essentially amorphous. The X-ray diffraction patterns of $Cu_{0.2}MnO_x$ and $Cu_{0.3}MnO_x$ have been plotted against temperature and presented in Fig. S3. As shown in Fig. S3, both materials heated in N_2 up to 400°C still remained amorphous. For $Cu_{0.2}MnO_x$ (Fig. S4),

it is revealed that a predominant phase of Mn_3O_4 and a minor phase of spinel $CuxMn_{3-x}O_4$ can be observed when temperature heated up to 500°C, and the additional peak of $CuMnO_2$ oxide (crednerite) can be found from 600 to 800°C. While, XRD of $Cu_{0.3}MnO_x$ showed that $CuMn_2O_4$ (PDF 34-1400) spinel oxide was formed and kept thermally stable in that temperature range of 500 to 800°C. The amorphous structure promotes the occurrence of lattice defects and the exposure of internal atoms, thus increasing the surface area of the catalyst [25], as seen in Table 1, the surface area of $Cu_{0.2}MnOx$ (154.3 m²/g) and $Cu_{0.3}MnO_x$ (156.3 m²/g) are much larger than spinel structure of $CuMn_2O_4$ (61.4 m²/g).

The morphologies of catalysts were investigated by SEM and HRTEM. Cu_{0.2}MnO_x and CuMn₂O₄ both show a similar morphology that aggregates nano-particles to form bigger particles. $CuMn_2O_4$ has particle size in the range of 20-40 nm, while, Cu_{0.2}MnO_x sample exhibits smaller size of spherical particles in the range of 8-15 nm, and the $Cu_{0.3}MnO_x$ sample shows the surface morphology of a rod shape with the length of ~360 nm. The inset in Fig. 1A2, B2, C2 shows the selected area electron diffraction (SAED) pattern, and the morphologies of catalysts exhibit a polycrystalline structure in all samples. Furthermore, as shown in Fig. 1A3, it can be seen that there are interplanar spacing of 0.305 and 0.159 nm in CuMn₂O₄, corresponding to 220, 511 crystal planes, respectively, and the angle of the 220 and 511 crystal planes are 35°. The fast Fourier transform (FFT) index of CuMn₂O₄ was [1,1,-4] zone axis, indicating the crystal plane of {-114} was selectively exposed. Comparing to amorphous samples, regular and smooth lattice fringes of CuMn₂O₄ suggested a good crystallization as expected. On the other hand, it can be observed that sporadic crystals are randomly distributed on the "amorphous" matrix (highlighted by dark yellow ovals) for Cu_{0.2}MnO_x. It was proposed the lattice fringes are covered by a large number of point defects [26], and these defects were believed to be the oxygen vacancies. It is reported that the oxygen vacancy normally is located in the second layer of catalyst crystal structure [32]. Therefore, when oxygen defect exists in the structure, the adjacent metal atom will permeate to the defect location, resulting in a shaded region, and the shaded region was referred to as oxygen vacancy on the catalyst's surface as seen in blue ovals in Fig. S5. As for the existence of sporadic crystal domains, the minimum lattice spacing of Cu_{0.2}MnO_x and Cu_{0.3}MnO_x is 0.162 nm and 0.185 nm, respectively.

The information of chemical composition was investigated using XPS, the relevant results were illustrated in Fig. 2 and Table 1. Cu 2p XPS spectra were characterized by two distinct regions of 947-957 eV and 930-938 eV that corresponding to Cu $2p_{1/2}$ and Cu $2p_{3/2}$ spin orbitals, respectively. The peaks at 930.6 and 950.5 eV can be attributed to Cu⁺. The signal at 933.4 and 953.2 eV accompanying with two shake-up satellites (S1 and S2) is Cu^{2+} [20]. Note that Cu^{+} was detected only on the surface of CuMn₂O₄. The spectrum of Mn 2p can be deconvoluted into 6 peaks. The peaks located at 641.1 and 652.5 eV were referred to Mn³⁺, and the signal at 639.7 and 647.5 eV were indicative of Mn²⁺ [27], whereas the Mn 2p signals at 643.0 and 655.8 eV were for Mn⁴⁺ [28]. The average oxidation state of Mn (Mn AOS) was calculated by the binding energy (ΔE) difference between the two peaks in Mn 3s (Mn AOS = 8.956 - 1.126 × Δ E) [29]. The results showed that the order of Mn AOS was CuMn₂O⁴ (3.8) > $Cu_{0.3}MnO_x$ (3.6) > $Cu_{0.2}MnO_x$ (3.5), which indicates the possible form of Mn^{3+} or/and Mn^{2+} in samples, and the content of Mn⁴⁺ was higher in the spinel sample comparing to amorphous samples. Moreover, it was reported the bonds of Mn²⁺-O and Mn³⁺-O are weaker than Mn⁴⁺-O [18], and more existence of low-valence Mn can weaken the Mn-O bond on the surface of the sample promoting the dissociation and activation of surrounding oxygen atoms. The atomic Cu/Mn ratio of CuMn₂O₄, CuO.₂MnO_x, and Cu_{0.3}MnO_x from the results of energy dispersive X-ray spectroscopy (EDX) was 0.5, 0.6 and 2.1 comparing to 0.5, 0.2 and 0.3 from ICP-OES results, respectively, revealing more Mn on the amorphous catalysts surface.

The O 1s spectra can be deconvoluted into four peaks, corresponding to surface lattice oxygen (Olat, 529.8 eV), adsorbed oxygen (O_{ads}, 531.2 eV), vacancy oxygen (Ov,532.2 eV), and surface hydroxyl

oxygen (Sur-OH, 532.8 eV), respectively [30, 31]. Notably, the O_{lat} mainly plays an oxidizing role in the O_3 catalytic process, and is usually supplemented by the following path: O_3 (O^- , O_2^- , $O_2^{2^-}$) + $Ov \rightarrow O_{ads} \rightarrow O_{lat}$ [33]. Therefore, as shown in Fig. 3, a good performance of $Cu_{0.2}MnO_x$ could be achieved due to a higher O_{ads}/O_{lat} (0.8). The O_{ads}/O_{lat} of $CuMn_2O_4$ and $Cu_{0.3}MnO_x$ is 0.7 and 0.6, respectively.

The unpaired electrons in the samples were detected by EPR. As shown in Fig. 4, the signal of g value from the amorphous sample ($Cu_{0.2}MnO_x$ and $Cu_{0.3}MnO_x$) presents a slight reduction compared to that of the $CuMn_2O_4$ (1.9987). The results indicate that the enhancement of the electron density on the amorphous sample interface, contributing to the numerous lattice boundaries accompany with more oxygen vacancy [34]. It is expected that the electron-rich environment and the existence of more oxygen vacancies can effectively supplement the electron consumption of samples during catalytic ozonation.

Fig. 5 exhibits the samples of H₂-TPR profiles. Two reduction peaks can be observed for each sample on the profiles, suggesting a continuous reduction of Cu-Mn cations. $Cu_{0.2}MnO_x$ and $CuMn_2O_4$ show the first peaks of hydrogen consumption at 227°C and 229°C, respectively, which are lower than that (252°C) of $Cu_{0.3}MnO_x$. This suggests the redox reactions of homemade catalysts at low temperatures are more active than the commercial catalyst. The continuous peaks were thought to be the reduction of the mixed-valence state of the Cu-Mn oxides, and the existence of Mn_xO_y also can prevent the reduction by forming a relatively stable $_{MnOx}$, resulting in the tailing peaks [35].

3.2. Catalytic ozonation of toluene

The toluene removal efficiency of three catalysts in 130 min as shown in Fig. 3A, exhibiting the order as follows: $Cu_{0.2}MnO_x$ (91.2%) > $CuMn_2O_4$ (74.5%) > $Cu_{0.3}MnO_x$ (70.3%). No obvious decrease in toluene removal efficiency was observed using all samples. Meanwhile, as Fig. 3B shown, better performances of O₃ decomposition at 100 °C were observed using homemade catalysts, while, for $Cu_{0.3}MnO_x$, O₃ concentration decreased to zero at the beginning, and then bounced back to ~400 ppm. It seems that the low temperature limits the O₃ decomposition of $Cu_{0.3}MnO_x$, leading to low toluene abatement. In order to confirm the role of O₃, the degradation of toluene was conducted without O₃ as shown in Fig. S6. Toluene removal efficiency vibrated at ~3%, and no obvious CO₂ was revealed in experiments. This suggested the O₃ was vital in toluene abatement as reported [36]. According to an overview table containing Cu or Mn-based ozonation of toluene in Table S1, the higher O₃ abatement was observed in our work, and the toluene removal efficiency is comparable to other works at ~100°C. Furthermore, based on the result of carbon balance in 130 min, the higher CO₂ yield was observed using $Cu_{0.2}MnO_x$, and the yield of CO_x (CO₂ + CO) was higher than 80%. For the spinel $CuMn_2O_4$, CO₂ yield was lower than that of amorphous samples, and it was revealed that ~ 17.7% of toluene turned into organic carbonaceous by-products.

3.3. The analysis of toluene decomposition

The toluene degradation pathway was explored through the technologies of GC-MS, DRIFTS, and TOF-SIMS. The results were displayed in Fig. 6-8 and Table S2. As shown in Fig. 6, a series of typical by-products (such as benzyl alcohol, benzaldehyde, benzoic acid, and maleic anhydride) were produced during the catalytic ozonation of toluene as reported [11, 37]. Note that other byproducts can be sorted into two types: benzenoids and ring-opening organic compounds (Table S3). It was revealed that the IR band at 1654 cm⁻¹ was presented, which was assigned to the C=O stretching vibration of quinone compounds, the results in GC-MS spectra also confirmed the presence of O-methyl benzoquinone. The possible reaction route ③ was proposed as shown in Fig. 8 (C₇H₈ \rightarrow C₆H₄ (CH₃)-OH \rightarrow HO-C₆H₃(CH₃)-OH \rightarrow O=C₆H₃(CH₃)=O) [11]. The strong peaks of 1450 and 1555 cm⁻¹ were assigned

to the antisymmetric and symmetrical stretching vibration of COO-. The peaks at 1410, 1498, and 1600 cm⁻¹ were considered to be in-plane skeletal vibrations of the aromatic ring, hence, the typical vibrations of aromatic ring and COO- were ascribed to the typical peaks of benzoic acid. Note that the COO- vibration of benzoic acid for home-made catalysts was much weaker than that with the use of Cu_{0.3}MnO_x, and it was observed the COO- peaks of Cu_{0.3}MnO_x sample weakened after 5 min reactions. Moreover, the C-OH stretching vibration of alcohols (1039 cm⁻¹) was discovered for all samples, indicating the presence of phenyl methanol. The above discussed byproducts were also confirmed by GC-MS results. Note that propanol was found in GC-MS spectra, which may be further oxidized to propionic acid [38] as Eq. 1.1-1.2 in Table 2. The appearance band at 1394 cm⁻¹ was referred to the - CH2 deformation vibration on the CuMn₂O₄ sample, which is often associated with carboxy (-CH₂-COOH). However, the -CH₂ vibration was not found in DRIFTS spectra of amorphous samples. The peaks at 2340 and 2361 cm⁻¹ were attributed to the generation of CO₂, and the peak at 2189 cm⁻¹ was assigned to CO. No obvious characteristic peaks of toluene were observed (at 3072 cm-1), it seems that the oxidation of toluene to the mentioned intermediates given above was faster than the adsorption of toluene.

In addition, the detection of $Cu_xMn_yO_zH_w^+$ secondary ions using TOF-SIMS has been depicted in Fig. S7. The CuOMn+ secondary signal contributes in relative intensity to 1.9%, 3.7%, and 1.6% in polarity (+) based on the total intensity of CuMn₂O₄, Cu_{0.2}MnO_x, Cu_{0.3}MnO_x, respectively. This indicates the better Mn-O-Cu interactions for Cu_{0.2}MnO_x. In addition, $C_7H_7^+$ (toluene) appeared as seen in Fig. 9, which was thought to be adsorbed firstly. C₇OH₅⁺ (benzyl alcohol), C₇OH₇⁺ (benzaldehyde), and C₇O₂H₅⁺ (benzoic acid) were also revealed as the results of DRIFTS, and were considered to be intermediates in the decomposition of toluene to benzoic acid (Eq.1.3-1.6). In addition, some ions, such as $C_6H_5O^+$ and $C_7O_2H_7^+$ referred to phenol and O-methyl hydroquinone, were involved in the formation of ptoluquinone $(C_7O_2H_7^+)$ in agreement with the previous discussion. Some other ring-breaking compounds such as maleic anhydride ($C_4O_3H^+$), acetic acid ($C_2H_3O_2^+$), and formic acid (HCO_2^+) were revealed, involving the formation of benzyl formate and benzyl acetate. In short, three possible toluene oxidation pathways were proposed based on the above analysis (Fig. 8). One is that toluene was oxidized for the formation of benzaldehyde, and then further oxidized to form benzoic acid. Hence, benzene could be formed through the decarboxylation process or demethylation of toluene. For ringopen reactions, the maleic anhydride, propanol, or other acids (acetic acid, formic acid) were generated as main intermediate products prior to the production of CO₂ or CO. No obvious diversity of byproducts among the three samples, but the intermediate products were oxidized in different reaction rate, leading to the diversity of carbon balance, which was in agreement with different intensity of IR bands in different samples.

3.4. The mechanism analysis of catalytic ozonation

The characteristics of post-experiment samples were analyzed, as seen in Fig. S8, no obvious phase change was observed. The morphologies of samples showed grain aggregation as seen in Fig. S9. This was also confirmed by pore diameter results in Table 1. Dp increased after the ozonation process. In addition, it was suggested that the carbon deposit may cause the change of SBET [39], the decrease of SBET was observed for all samples. Furthermore, TGA/DTG curves for samples are displayed in Fig. S10 and summarized in Table S4. It was observed that a total weight loss of 6.7, 15.8, and 23.7 wt% for CuMn₂O₄, Cu_{0.2}MnO_x, and Cu_{0.3}MnO_x, respectively. According to DTG curves, the weight losses from 30 to 300°C were ascribed to water desorption (30-150°C) and chemically bonding water (150-300°C). The second weight loss of 1.4 wt%, 3.7 wt%, and 5.5 wt% at 300-600°C was found for CuMn₂O₄, Cu_{0.2}MnO_x, and Cu_{0.3}MnO_x, respectively, suggesting the desorption of carbon deposits [40].

Commercial catalyst $Cu_{0.3}MnO_x$ showed the highest weight loss, suggesting the incomplete ozonation process.

The valence states of samples on the surface were characterized using XPS (Fig. S11). After ozonation reactions, the Cu⁺ content of CuMn₂O₄ (27.5%) and Mn AOS (3.8) decreased to 13.1% and 3.6, respectively. The possible reaction was proposed: $\equiv Mn^{4+} + \equiv Cu^{2+} + 2O_{lat} \rightarrow \equiv Mn^{2+}/Mn^{3+} + Vo + \equiv Cu^{+} + O_2$ [10], indicating that Cu-O (Cu⁺/Cu²⁺) exist as an electron mediator on the degradation of toluene, while, it may as an adsorption site, giving electrons for O₃ decomposition to form active O species, then, toluene was further oxidized on the site of Mn⁴⁺ as Fig. 10 shown. The bonding of Cu⁺-O²⁻-Mn⁴⁺ of the spinel structure, exhibiting highly oxidizing ability, transformed to the Cu²⁺-O²⁻-Mn³⁺ in the oxidation of toluene [41].

On the other hand, Mn AOS of Cu_{0.2}MnO_x and Cu_{0.3}MnO_x decreased from 3.5 and 3.8 to 3.0 and 3.6, respectively, indicating the reduction of Mn. It is worth noting that Cu-O (Cu⁺/Cu²⁺) generally participates in the reaction as an adsorption site [42], and the valence state of Cu was unchanged in the amorphous sample, indicating the main ozonation process occurred via Mn atom (Eq. 2.1-2.4). This may suggest the different ozonation processes between amorphous samples and spinel samples, i.e. the possible replenishment process of electrons on Cu⁺/Cu²⁺ in amorphous samples. It was observed the ratio of O_{ads}/O_{lat} (Table 1) increased after reaction implied the consumption of O_{lat} , which was in agreement with the increase of O_v for amorphous samples. Therefore, the structural defects of amorphous cannot be ignored on the surface of samples. Hence, it is suggested the strong oxidizing group (O⁻, O₂⁻, O₂²⁻) produced by the decomposition of O₃ could be captured by Vo due to the positive property of Vo, and then the adjacent adsorbed toluene was oxidized by reactive O to form the byproducts as discussed above.

All in all, according to the above analyses, the mechanism of catalytic ozonation of toluene could be proposed as presented in Fig. 10. At first, toluene was adsorbed into Cu^+/Cu^{2+} (Eq. 3.1-3.2). For amorphous samples, toluene was chemisorbed to be C_7H_7 - by the electrons from charge-transfer of $Cu^+ \rightarrow Cu^{2+}$, the replenishment of $Cu^{2+} \rightarrow Cu^+$ occurred due to the formation of oxygen vacancies. These oxygen vacancies were thought to be generated from the process of $Mn^{4+} \rightarrow Mn^{2+}/Mn^{3+}$ in accompany with the decomposition of O_3 , leading to the decrease of Mn AOS (Eq. 2.1-2.3). Hence, Mn was believed to be the most reactive site on the amorphous sample for catalytic ozonation. On the other hand, for $CuMn_2O_4$, the electrons from $Cu^{2+} + Mn^{2+}/Mn^{3+} \leftrightarrows Cu^+ + Mn^{3+}/Mn^{4+}$ were used for ozonation process. Considering to generation of Vo during O3 decomposition, the chemical reaction equilibrium of Eq. 4.1 shifts to left, leading to a lower content of Cu^+/Mn^{4+} .

4. Conclusion

The catalytic activities of various Cu-Mn catalysts (spinel and amorphous) for the oxidation of toluene have been examined. Firstly, it is observed that the catalytic ozonation of toluene has higher efficiency (91.2%) and CO_x yield (83.0%) by using Cu_{0.2}MnO_x at 130 min in comparison with other samples. Secondly, the chemisorption and catalytic capabilities of Cu-Mn catalysts can be attributed to the synergistic effects between Cu and Mn, such as the electron-transfer process. The well-dispersed Vo on the surface of amorphous catalysts helped the chemisorption of O₃, and adjacent Mn was believed to be as a main active site in the reaction. In addition, CuMn₂O₄ exhibited spinel crystal structure, adjunction sites of Cu⁺/Mn⁴⁺ were considered to be reactive sites. The consumed redox pair inhibited toluene further abatement. At last, the possible routes of ozone-assisted catalytic oxidation toluene were proposed over Cu-Mn catalysts, the different physic-chemical properties of Cu-Mn catalysts result in different reaction rates of intermediates decomposition, leading to different amounts of byproducts and toluene abatement efficiency.

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Figures:



Fig. 1. SEM images of various Cu-Mn samples of $CuMn_2O_4$ (A1), $Cu_{0.2}MnO_x$ (B1), and $Cu_{0.3}MnO_x$ (C1) catalysts; HRTEM images of $CuMn_2O_4$ (A2-3), $Cu_{0.2}MnO_x$ (B2-3), and $Cu_{0.3}MnO_x$ (C2-3) catalysts.



Fig. 2. XPS spectra of CuMn₂O₄, Cu_{0.2}MnO_x, and Cu_{0.3}MnO_x. (A) Cu 2p, (B) Mn 2p, (C) O 1s.



Fig. 3. Toluene abatements on various samples at 100° C, ~160 ppm toluene, and ~2100 ppm O₃. (A)Toluene removal, (B) O3 residual, (C) Carbon balance at 130 min.



Fig. 4. EPR spectra of various samples



Fig. 5. H₂-TPR profiles of various samples.



Fig. 6. GC-MS spectra of gaseous by-products by using various samples in the ozone-catalysis system (Lines a-c in order: CuMn₂O₄, Cu_{0.2}MnO_x, Cu_{0.3}MnO_x): (1) Propanol, (2) Toluene, (3) Maleic anhydride, (4) Benzaldehyde, (5) O-methyl benzoquinone, (6) Benzyl alcohol, (7) Benzyl formate, (8) Benzoic acid, (9) Benzyl acetate.



Fig. 7. DRIFTS of the substances on the surface of samples at 100° C and 2100 ppm O_3 . (A) CuMn₂O₄ (Lines a-d in order: 1 min, 5 min, 16 min, 130 min), (B) CuO.2MnOx (Lines a-e in order: 1 min, 3 min, 5 min, 16 min, 130 min), (C) Cu_{0.3}MnO_x (Lines a-d in order: 1 min, 3 min, 5 min, 16 min, 130 min).



Fig. 8. The proposal reaction pathway over Cu-Mn catalyst for toluene oxidation with the presence of O_{3} .



Fig. 9. The $C_xO_yH_z$ ions in polarity (+) for three samples: (1) HCO_2^+ , (2) $C_3H_7O^+$, (3) $C_4O_3H^+$, (4) $C_6H_5^+$, (5) $C_6H_5O^+$, (6) $C_7H_7^+$, (7) $C_7OH_5^+$, (8) $C_7OH_7^+$, (9) $C_7O_2H_5^+$, (10) $C_7O_2H_7^+$.



Fig. 10. The mechanisms of Cu-Mn bimetallic sites synergetic ozone-catalyzed oxidation of toluene.

Table 1 The relevant data about fresh and post reaction catalysts.

Table 1

The relevant data about fresh and post reaction catalysts.

Samples		S _{BET} ^a (m ² /g)	V _p ^b (cm ³ / g)	D _p ° (nm)	O _{ads} / O _{lat} ^d	Cu ⁺ / (Cu ⁺ +Cu ²⁺) ^d	$Cu^{2+}/(Cu^{+}+Cu^{2+})^{d}$	AOS _{Mn} ^d	Cu∕ Mn ^e	Cu∕ Mn ^f	H ₂ -TPR ^g (°C)
Fresh	CuMn ₂ O ₄	61.4	0.1	4.7	0.7	27.5	72.5	3.6	0.5	0.5	275
	CuMn _{0.2} O _x	154.3	0.5	5.6	0.8	-	100	3.5	0.2	0.6	263
	CuMn _{0.3} O _x	156.3	0.2	2.3	0.6	-	100	3.8	0.3	2.1	283
Post-	CuMn ₂ O ₄	30.1	0.1	14.8	0.7	13.1	86.9	3.2			
reaction	CuMn _{0.2} O _x	105.0	0.3	10.1	1.0	-	100	3.0			
	CuMn _{0.3} O _x	129.0	0.2	5.5	0.8	-	100	3.6			

^a Surface Area.

^b Desorption cumulative volume of pores.

^c Desorption average pore diameter. ^d From XPS results.

^e Cu/Mn of the entire catalyst obtained from ICP-OES results ^f Cu/Mn of the entire catalyst obtained from EDX results.

^g The maximum temperature of reduction.

Table 2: Possible reactions for ozone catalytic oxidation process.

Catalytic ozonation process	
$CH_3CH_2CH_2-OH + O_{lat} \rightarrow CH_3CH_2CHO$	Eq. 1.1
$CH_3CH_2CHO + O_{lat} \rightarrow CH_3CH_2COOH$	Eq. 1.2
$C_6H_5-CH_3 + Ov \rightarrow C_6H_5-CH_3-O_v$	Eq. 1.3
C_6H_5 - CH_3 - $Ov + O_{lat} \rightarrow C_6H_5$ - CH_2OH	Eq. 1.4
C_6H_5 - $CH_2OH + O_{lat} \rightarrow C_6H_5$ - $CHO + H_2O$	Eq. 1.5
C_6H_5 -CHO + $O_{lat} \rightarrow C_6H_5$ -COOH	Eq. 1.6
Mn element reaction process	
$Mn^{2+}/Mn^{3+} + O_3 + O_v \rightarrow O_2 + O_2^{2-}ads + Mn^{3+}/Mn^{4+}$	Eq. 2.1
$O_2^{2^-}_{ads} + Mn^{3+}/Mn^{4+} \rightarrow Mn^{3+}/Mn^{4+} + O^{2-} + O_{ads}$	Eq. 2.2
$Mn^{3+}/Mn^{4+} + O_{lat} \rightarrow Mn^{2+}/Mn^{3+} + O_2 + O_v$	Eq. 2.3
$C_7H_8 + Mn^{3+}/Mn^{4+} + O_{ads}/O_{lat} \rightarrow H_2O + CO_2 + Mn^{2+}/Mn^{3+} + Ov$	Eq. 2.4
Toluene chemisorption	
$Cu^+ + C_7H_8 \rightarrow C_7H_7$ - $Cu + H^+$	Eq. 3.1
Cu ⁺ + 2 C ₇ H ₈ → C ₇ H ₇ -Cu-C ₇ H ₇ + 2 H ⁺	Eq. 3.2
Charge transfer process	
$Cu^{2+} + Mn^{2+}/Mn^{3+} \Leftrightarrow Cu^{+} + Mn^{3+}/Mn^{4+}$	Eq. 4.1
$Cu^{+} + Mn^{4+}/Mn^{3+} \leftrightarrows Mn^{2+}/Mn^{3+} + Cu^{2+}$	Eq.4.2

Supporting Information

Investigation of Cu-Mn catalytic ozonation of toluene: Crystal phase, intermediates and mechanism

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Text

1. Relevant calculation

Figure and Table

Fig. S1. Schematic diagram of catalyst activity testing device.

Fig. S2. Powder XRD patterns of CuMn₂O₄, Cu_{0.2}MnO_x, and Cu_{0.3}MnO_x.

Fig. S3. In situ XRD patterns of $Cu_{0.3}MnO_x$ under flowing N_2 as a function of temperature.

Fig. S4. In situ XRD patterns of $Cu_{0,2}MnO_x$ under flowing N2 as a function of temperature.

Fig. S5 HRTEM images of CuMn₂O₄ (A), Cu_{0.2}MnO_x (B), and Cu_{0.3}MnO_x (C) catalysts.

Fig. S6. Toluene abatements on various samples without O_3 . (A) Toluene removal rate, (B) CO_2 yield.

Fig. S7. The $Cu_xMn_yO_zH_w$ ions in polarity (+) for three samples: (1) MnCuO, (2) ^{^65}CuMnO, (3) CuMnOH, (4) ^{^65}CuMnOH, (5) CuMn, (6) CuMnH, (7) CuMnO₂, (8) CuMnO₂H⁺, (9) CuMnO₂H₂, (10) ^{^65}CuMnO₂H, (11) ^{^65}CuMnO₂H₂, (12) CuMn₂O₂, (13) ^{^65}CuMn₂O₂, (14) CuMn₂O₃, (15) ^{^65}CuMn₂O₃, (16) CuMn₂O₃H, (17) ^{^65}CuMn₂O₃H, (18) CuMn₃O₃, (19) ^{^65}CuMn₃O₄, (21) ^{^65}CuMn₃O₄.

Fig. S8. Powder XRD patterns of the post-reaction various Cu-Mn samples.

Fig. S9. SEM images of the post-reaction Cu-Mn samples. (A) $CuMn_2O_4$, (B) $Cu_{0.2}MnO_x$, (C) $Cu_{0.3}MnO_x$.

Fig. S10. TGA/DTG analyses of the post-reaction catalysts.

Fig. S11. XPS spectra of the post-reaction various Cu-Mn samples. (A) Cu 2p, (B) Mn 2p, (C) O 1s.

Table S1 Catalytic ozonation of toluene by various samples.

Table S2 Attribution of main characteristic peaks in DRIFTS.

Table S3 Main by-products in the exhaust gas.

Table S4 TG of various samples.

1. Relevant calculation

Toluene removal rate (TRR), CO yield (X_{CO}) and CO₂ yield (X_{CO2}), and others (X_0) were calculated as follows:

$$TPR (\%) = \frac{[Toluene]_{inlet} - [Toluene]_{outlet}}{[Toluene]_{inlet}} \times 100\%$$
(1)

$$X_{CO} (\%) = \frac{[CO]}{[Toluene]_{inle} \times 7} \times 100\%$$

$$\tag{2}$$

$$Xco_{2} (\%) = \frac{[CO_{2}]}{[Toluene]_{inlet} \times 7} \times 100\%$$
(3)

$$X_{O}$$
 (%) = (1 - X_{CO} - X_{CO_2} - TPR) × 100% (4)

Here, [Toluene]_{inlet} is toluene concentration (ppm) in the mixture at the inlet, [Toluene]_{outlet} is toluene concentration (ppm) in the mixture at the outlet. [CO] and [CO₂] are the concentration of CO_2 and CO (ppm) in the gas stream outlet during the oxidation reaction, respectively.



Fig. S1. Schematic diagram of catalyst activity testing device.



Fig. S2. Powder XRD patterns of fresh samples of CuMn₂O₄, Cu_{0.2}MnO_x, and Cu_{0.3}MnO_x.



Fig. S3. In situ XRD patterns of $Cu_{0.3}MnO_x$ under flowing N_2 as a function of temperature.



Fig. S4. In situ XRD patterns of $Cu_{0.2}MnO_x$ under flowing N_2 as a function of temperature.



Fig. S5. HRTEM images of CuMn₂O₄ (A), Cu_{0.2}MnO_x (B), and Cu_{0.3}MnO_x (C) catalysts.



Fig. S6. Toluene abatements on various samples without O_3 . (A) Toluene removal rate, (B) CO_2 yield.



Fig. S7. The $Cu_xMn_yO_zH_w$ ions in polarity (+) for three samples: (1) MnCuO, (2) ⁶⁵CuMnO, (3) CuMnOH, (4) ⁶⁵CuMnOH, (5) CuMn, (6) CuMnH, (7) CuMnO₂, (8) CuMnO₂H⁺, (9) CuMnO₂H₂, (10) ⁶⁵CuMnO₂H, (11) ⁶⁵CuMnO₂H₂, (12) CuMn₂O₂, (13) ⁶⁵CuMn₂O₂, (14) CuMn₂O₃, (15) ⁶⁵CuMn₂O₃, (16) CuMn₂O₃H, (17) ⁶⁵CuMn₂O₃H, (18) CuMn₃O₃, (19) ⁶⁵CuMn₃O₃, (20) CuMn₃O₄, (21) ⁶⁶⁵CuMn₃O₄.



Fig. S8. Powder XRD patterns of the post-reaction various Cu-Mn samples.



Fig. S9. SEM images of the post-reaction Cu-Mn samples. (A) $CuMn_2O_4$, (B) $Cu_{0.2}MnO_x$, (C) $Cu_{0.3}MnO_x$.



Fig. S10. TGA/DTG analyses of the post-reaction catalysts.



Fig. S11. XPS spectra of the post-reaction various Cu-Mn samples. (A) Cu 2p, (B) Mn 2p, (C) O

1s.

Samples	Metal loading (wt%)	Temperature (°C)	O ₃ (ppm)	Toluene (ppm)	Reaction time (min)	Toluene removal (%)	O3 removal (%)	Ref.
Cu (imp)/MCM-41	4	22	1345	395	88	33	38	[1]
Cu (insitu)/MCM- 41	4	22	1345	395	88	15	36	[1]
δ -MnO ₂ /USY	3	25	60	20	180	55	100	[2]
Cu-Mn/HZSM-5	1(Cu), 5(Mn)	25	1000	100	-	27	27	[3]
Ru-Mn/HY (80)	1(Ru), 5(Mn)	25	1000	100	-	63	89	[4]
Cu/13X	2	90	7650	896	100	15	24	[5]
Mn/13X	2	90	7650	896	100	47	63	[5]
MnO _x /γ-Al ₂ O ₃	5	120	1000	100	150	100	100	[6]
$Cu_{0.2}MnO_x$	-	100	2100	160	130	91	100	
CuMn ₂ O ₄	-	100	2100	160	130	75	100	
$Cu_{0.3}MnO_x$	-	100	2100	160	130	70	80	

Table S1 Catalytic ozonation of toluene by various samples.

Table S2 Attribution of main characteristic peaks in DRIFTS.

Wavenumber (cm ⁻¹)	Ascription	Characteristic substances
1026, 1056, 2122		O ₃
1039	C-OH stretching vibration	Alcohols
1180, 1275	C-O stretching vibration	Alcohols or phenols
1309	C-OH Vibration peak	Phenols
1368	-CH ₃ Vibration peak	Methyl
1394	-CH ₂ Vibration peak	Methylene
1410, 1498, 1600	C=C Vibration peak	Benzene ring
1450, 1555	COO- stretching vibration	Acids
1656, 1696, 1755, 1785	C=O stretching vibration	Quinone, aldehydes, acids
		or esters
2195	CO	CO
2340, 2361	CO ₂	CO_2
3072	C-H aromatic ring	Toluene

Number	Peak time	Chemical compound	Molecular	Structural formula
	(min)		formula	
D	5.414	Propanol	C_3H_8O	HO
organic compound	10.591	Maleic anhydride	$C_4H_2O_3$	
	8.005	Toluene	$\mathrm{C_{7}H_{8}}$	
	14.255	Benzaldehyde	C7H6O	
	16.031	O-methylbenzoquinone	$C_7H_6O_2$	
Benzenoids	16.709	Benzyl alcohol	C_7H_8O	ОН
Denzenoles	18.18	Benzyl formate	$C_8H_8O_2$	
	20.646	Benzoic acid	$C_7H_6O_2$	ОН
	20.925	Benzyl acetate	$C_{9}H_{10}O_{2}$	

Table S3 Main by-products in the exhaust gas.

Table S4 TG of various samples.

Samples	30-300°C (%)	300-600°C (%)	600-790°C (%)	30-790°C (%)
CuMn ₂ O ₄	4.5	1.4	0.8	6.7
$Cu_{0.2}MnO_x$	11.7	3.7	0.4	15.8
$Cu_{0.3}MnO_x$	16.5	5.5	1.7	23.7

Reference

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CrediT authorship contribution statement

Zhiping Ye: validation, results analysis, methodology, and writing; Guanjie Wang: Test experiment, results analysis, methodology, and writing; Jean-marc Giraudon: Conceptualization, validation, writing-review and editing, Anton Nikiforov: Conceptualization, validation, writing-review and editing, Jun Chen: Conceptualization, validation, writing-review and editing; Liang Zhao: Catalyst characterization; Xiuwen Zhang: Catalyst characterization; Jiade Wang: Funding support, supervision, validation.

Declaration of Competing Interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.



Graphical abstract

Highlights

• Amorphous phase of Cu-Mn oxides is more active than spinel phase of sample in catalytic ozonation of toluene.

• High dispersion of Mn and oxygen vacancy played important role for amorphous samples in catalytic ozonation process.

• Redox pair of Cu⁺/²⁺/Mn⁴⁺/3⁺ in spinel acted as active site for

degradation of ozone and toluene.

• No obvious difference was observed in the distribution of byproducts, except for the quantities.